

Persistent winter nitrate pollution driven by increased oxidants in northern China



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Chinese Research Academy of Environmental Sciences

Atmospheric Chemistry Mechanisms Conference Organized by UC-Davis, Virtual Conference 9-19 November 2020



Acknowledgement

- Prof. Shuxiao Wang of Tsinghua University
- Prof. Likun Xue of Shandong University
- National Natural Science Foundation of China (91544213)
- Hong Kong Research Grant Council (ANR-RGC Joint research and Theme-based research T24-504/17-N)



- Winer haze in the North China Plain (NCP)
- Nitrate observations in Dec 2017 in NCP
- Key factors controlling nitrate formation and trend (by model)

The North China Plain (NCP)



NOx emission



~0.3 million km² and ~1/5 Chinese population

Home to Beijing, Tianjin, Shandong, and part of Hebei, Henan, Jiangsu and Anhui



$SO_2 \& NO_x$ emissions decreased, so did ambient $PM_{2,5}$ conc.

-17%

NO_x

2014

China emission changes (2010-2017)







-Δ







But no obvious decline in fine nitrate in NCP



The trends were composed using the results from Zhao et al., 2019; Meng, 2015; Ma, 2017; Wang et al., 2019; Jia et al., 2018; Zhang et al., 2017; Jia et al., 2018; Wen et al., 2016; Han et al., 2015; Shao et al., 2018.

Nitrate formation pathways



Three key processes: HNO_3 production by $OH+NO_2$ and N_2O_5 hydrolysis followed by reaction with NH_3 **Three key ingredients**: NO_x , NH_3 &oxidants

Evidence of active winter photochemistry

Peak concentrations of HONO, PAN, and OH observed during wintertime in the NCP



Regional observations of PM2.5 in Dec 2017



Highest pollution occurs south of Beijing

- % of nitrate increases in heavy pollution
- Nitrate is more than sulfate

Model simulations of winter nitrate in Dec 2017

• Model simulation

- CMAQ5.1 / WRF4.0
- Domain: 36km, 12km
- Period: 1-31 Dec. 2017

- CMAQ: SAPRC07tic + AERO6i
 - WRF: Pleim-Xiu + ACM2 + RRTMG
 The first guess fields: ds083.2 from NCEP
 - Grid nudging: ds351.0 and ds461.0
- Emission: Tsinghua + MEGAN



Improving the model on reactive nitrogen chem.

HONO sources (Fu et al., 2019)



 $k_4/k_{2b} = 2.9 \pm 1.2$

Location	Obs. period	Obs. average	Sim. average	Reference	_		$\gamma(N_2O_5)$	
ICCAS_Beijing	2014.12	1.34	2.40	Tong et al. (2016)				
CEE_Beijing	2016.01	1.05	2.40	Wang et al. (2017)		"Observed"	0.024 ± 0.023	
EPA_Beijing	2015.02-03	1.99	2.40	Zhang et al. (2018)		B&T nara	0.046 ± 0.015	
Jinan_Shandong	2016.12- 2017.02	1.75	1.92	Li et al. (2018)		Det para.	0.040±0.015	
Wangdu_Hebei	2017.12	2.27	0.81	unpublished data		Modified	0.023 ± 0.020	

 $k_4/k_{2b} = 29 \pm 6$

Improved simulations of NO₂ and nitrate in NCP



		OBS (µg m ⁻³)	SIM (μg m ⁻³)	Bias (µg m ⁻³)	NMB (%)	NME (%)	R
NO ₃ -	CAMQ default (BT09)	20.04	24.86	3.92	18.72	47.80	0.75
	CAMQ revised (Fitted)	20.94	20.98	0.04	0.19	41.70	0.75
NO ₂	CAMQ default (BT09)	52.00	45.71	-6.38	-12.25	41.87	0.56
	CAMQ revised (Fitted)	52.09	47.89	-4.20	-8.06	41.33	0.58

NCP is a NH₃-rich environment in winter



$$GR = \frac{\left(\left[NH_{3}\right] + \left[NH_{4}^{+}\right]\right) - 2 \times \left[SO_{4}^{2-}\right]}{\left[NO_{3}^{-}\right] + \left[HNO_{3}\right]}$$

GR>1 indicates NH₃-rich conditions 0<GR<1 indicates NH₃-neutral conditions GR<0 indicates NH₃-poor conditions

HNO₃ sources and sinks (region average)



Physical & chemical processes

Downward contribution to surface nitrate

Comparable gas-phase and het. reaction, higher $OH+NO_2$ at surface

Increased production of oxidants and HNO₃ in heavy pollution



- Due to increase in HONO, OVOCs, NOx, RH, despite ~30% reduction in sunlight
- Increased N₂O₅ hydrolysis in heavy haze



Why no improvement of nitrate from 2010 to 2017?

Emission changes from 2010 to 2017 in the NCP: SO₂ (-59.7%), NO_X (-31.8%), PM_{2.5} (-38.6%), VOC (4%), NH₃ (0.2%)





VOC control would have reduced nitrate

2010-2017 Emission change: $\rm NO_x$: -32%, $\rm SO_2$: -60%, $\rm PM_{2.5}$: -39% , VOCs: +4%, $\rm NH_3$: +0.2%



Conclusion

- Winter photochemistry in NCP is active enough to drive the formation of nitrate, due to high conc. of oxidant precursors (e.g. HONO, VOC).
- The PM targeting emission control measures in the past decade increased O_3 and OH, offsetting the effectiveness of NO_X reduction.
- Future strategies should also reduce the oxidants, via larger NO_X and VOC control.

View of Mt Tai in NCP

Thank you!



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For more info.

Article

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Cite This: Environ. Sci. Technol. 2020, 54, 3881-3889

